

## ADAPTATION AND EVOLUTIONARY RESPONSES TO HIGH CO<sub>2</sub>

PHILIP L. MUNDAY<sup>\*,†</sup>

JODIE L. RUMMER<sup>\*</sup>

HANNES BAUMANN<sup>†</sup>

\*ARC Centre of Excellence for Coral Reef Studies, James Cook University, Townsville, QLD, Australia

<sup>†</sup>Department of Marine Sciences, University of Connecticut, Groton, CT, United States

<sup>1</sup>Corresponding author: philip.munday@jcu.edu.au

1. Introduction
2. Evolutionary CO<sub>2</sub> History of Teleosts
3. Adaptation to High and Variable CO<sub>2</sub> Habitats
4. Phenotypic Plasticity and Genetic Adaptation to Future CO<sub>2</sub> Levels
  - 4.1. Within and Between Generation Plasticity to High CO<sub>2</sub>
  - 4.2. Heritability and Selection to High CO<sub>2</sub> in Fishes
  - 4.3. Selection Differentials
  - 4.4. Genetic Correlations
5. Selection in Aquaculture
6. Knowledge Gaps and Future Directions

### Acknowledgments

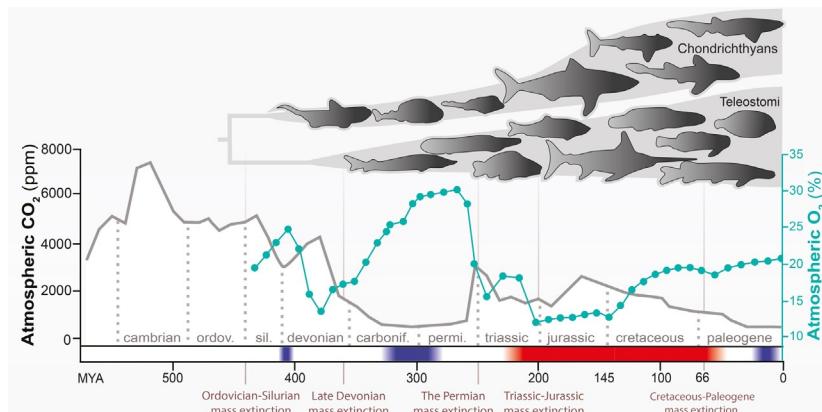
### References

The evolutionary history of fishes spans geological periods where atmospheric CO<sub>2</sub> was much higher than the current-day, yet some extant species are now sensitive to high environmental CO<sub>2</sub>. Other species have adapted to live in habitats where they naturally encounter very high CO<sub>2</sub> levels. This chapter explores the evolutionary history of fishes in relation to environmental CO<sub>2</sub> and adaptations to high CO<sub>2</sub> habitats. It then considers the potential for adaptive responses to predicted future CO<sub>2</sub> levels from climate change among extant fishes. Despite a rich theory and well-developed experimental methods in quantitative genetics only a handful of studies have tested for genetic

variation in CO<sub>2</sub>-sensitive traits, which might enable fish to adapt to projected future CO<sub>2</sub> levels. This is a serious knowledge gap that needs a concerted research effort to overcome. Without basic information on genetic variation in fitness-associated traits and the strength of selection, it is not possible to make informed decisions about the impacts of elevated CO<sub>2</sub> on fish populations over the timeframes that CO<sub>2</sub> is changing.

## 1. INTRODUCTION

Modern fishes began their evolutionary history during a geological period when atmospheric and aquatic CO<sub>2</sub> concentrations were much higher than today (Fig. 1, Clack, 2007; Rummer and Munday, 2017). This evolutionary history has likely helped to shape many of the physiological processes that define modern fishes, especially as they relate to respiration, acid-base regulation, and calcification (Randall *et al.*, 2014). Nevertheless, some extant fishes are sensitive to relatively small increases in environmental CO<sub>2</sub>, to levels that are still much lower than their ancestors would have experienced during their evolutionary history. Indeed, experiments over the past decade have revealed that some fish are sensitive to CO<sub>2</sub> levels predicted to occur in the ocean over the next few hundred years due to ongoing anthropogenic CO<sub>2</sub> emissions into the atmosphere (Cattano *et al.*, 2018). These experiments show that CO<sub>2</sub> levels between 1000 and 2000  $\mu\text{atm}$  can affect growth, development, swimming performance, survivorship, reproduction, and a broad range of ecologically



**Fig. 1.** Geological timeline documenting the evolutionary history of the chondrichthyans and Teleostomi with corresponding atmospheric CO<sub>2</sub> concentrations in parts per million (ppm, left y-axis), atmospheric O<sub>2</sub> percentage (right y-axis) and global warming and cooling periods noted by red and blue shaded bar (Clack, 2007; Erwin *et al.*, 2011; Kardong, 2012; Royer *et al.*, 2004).

relevant behaviors of some marine and freshwater fishes (Cattano et al., 2018; Heuer and Grosell, 2014; Nagelkerken and Munday, 2016; Chapter 9, Vol 37: Munday et al., 2019a). Two important conclusions can be reached from these observations. First, it is clear that many extant fishes are adapted to the low CO<sub>2</sub> levels that have dominated Earth's recent geological history (Masson-Delmotte et al., 2013; Rummer and Munday, 2017). While extant species may have experienced high CO<sub>2</sub> levels in their deep evolutionary history, this does not necessarily mean that they will tolerate high CO<sub>2</sub> levels in the future. Second, to reliably predict how fish populations will respond to higher aquatic CO<sub>2</sub> levels over the next 100–200 years due to accumulating anthropogenic CO<sub>2</sub> emissions, we must consider their potential to adapt to higher CO<sub>2</sub> levels over this timeframe (Munday et al., 2013b; Reusch, 2014; Sunday et al., 2014).

Atmospheric CO<sub>2</sub> concentrations have risen from approximately 280 ppm before the Industrial Revolution to over 400 ppm in the current day ([www.esrl.noaa.gov/gmd/ccgg/trends/global](http://www.esrl.noaa.gov/gmd/ccgg/trends/global)). While most of the ocean surface waters and large freshwater basins are in approximate gas equilibrium with the atmosphere (Doney et al., 2009), some habitats can have much higher PCO<sub>2</sub> (CO<sub>2</sub> partial pressure) due to a variety of physical and biological processes that concentrate CO<sub>2</sub> inputs (Fig. 2; Chapter 1, Vol 37: McNeil and Matsumoto, 2019). For example, coastal upwelling can bring CO<sub>2</sub>-rich water from the oxygen minimum zone to the surface in nearshore habitats (Feely et al., 2008). Furthermore, biological activity in shallow water habitats can enhance aquatic CO<sub>2</sub> content, especially when water flow is limited or constrained (Baumann and Smith, 2018). As an extreme example, small creeks and seasonal ponds that are connected to rivers can attain very high CO<sub>2</sub> tensions (>10,000 µatm) when water flows are restricted and/or CO<sub>2</sub> flux with the atmosphere is constrained by surface growing plants (Borges et al., 2015; Ultsch, 1996). Biological activity, and thus CO<sub>2</sub> production, can be further enhanced in freshwater and coastal habitats by the input of terrigenous nutrients. Despite periods of high CO<sub>2</sub>, these aquatic habitats can have diverse fish communities (Junk et al., 2007), demonstrating the capacity of fishes to adapt to habitats that naturally encounter periods of high CO<sub>2</sub>. In the most extreme cases, this has involved unique physiological and behavioral adaptations to deal with extended periods of hypoxia and hypercapnia. In contemporary times, the global expansion of aquaculture production has created another environment where fishes are often exposed to very high PCO<sub>2</sub>, due to the concentration of metabolic CO<sub>2</sub> in high-density culture (Chapter 8, Vol 37: Skov, 2019). Cultured fishes are also subjected to artificial selection and/or selective breeding of genotypes favoring high performance in aquaculture conditions, potentially also promoting high CO<sub>2</sub> tolerance, but at the expense of physiological and behavioral traits that may be needed to survive in the wild.



**Fig. 2.** A representative land-sea interface showing a variety of habitats that may experience high PCO<sub>2</sub> due to concentration of CO<sub>2</sub> inputs by biological and physical processes. Drawn by Erin Walsh.

There are two principal ways that animals can adjust to environmental change; they can acclimate through phenotypic plasticity or they can genetically adapt (evolution). Phenotypic plasticity is the capacity of a single genotype to produce different phenotypes depending on the environmental conditions experienced. Therefore, an individual may acclimate to a changing environment by expressing a phenotype that has higher performance under those particular conditions. By contrast, genetic adaptation involves selection of genotypes that have higher fitness under the new environmental conditions, leading to a change in the frequency of alleles in the population and a shift in the mean phenotype of the population toward the fitness peak (Hendry *et al.*, 2011). Importantly, phenotypic plasticity can act relatively rapidly, within an individual's life time; whereas genetic adaptation may take many generations, depending on the strength of selection and the amount of genetic variation in the population for the relevant traits (Chevin *et al.*, 2010; Crozier and Hutchings, 2014; Hoffmann and Sgro, 2011). Phenotypic plasticity and genetic adaptation may also influence each other. For example, within-generation phenotypic plasticity could slow the pace of among-generation genetic change by shifting the mean phenotype closer to

the fitness peak and thus reducing the strength of selection (Diamond and Martin, 2016). Furthermore, plasticity exhibited within a generation may be influenced by the environmental conditions experienced in previous generations (parental and transgenerational effects), meaning that plasticity can also be a multigenerational process (Bonduriansky et al., 2012) and can interact with genetic adaptation (Klironomos et al., 2013).

Predicting the adaptive potential of fish populations to future higher CO<sub>2</sub> levels is further complicated by the presence of additional environmental changes, such as global warming. There may be fitness trade-offs and/or genetic correlations between traits under elevated CO<sub>2</sub> and high temperature (or any other environmental stressor) that could either enhance or retard the pace of adaptation (Munday et al., 2013b; Sunday et al., 2014). At the same time, over-harvesting may reduce the genetic diversity of some fish populations and thus reduce their potential to respond to selection. Consequently, predicting whether fish will adapt to the combined effects of elevated CO<sub>2</sub> and other stressors is not an easy task.

In this chapter, we first explore the evolutionary history of modern fishes in relation to global trends in atmospheric CO<sub>2</sub> with a focus on the adaptive radiation of teleost fishes over the past 50 million years. We also consider extant fishes living in high CO<sub>2</sub> habitats and how they have adapted to these environments. We then examine what is known about the adaptive potential of fishes to rising CO<sub>2</sub> levels associated with global climate change. Laboratory experiments show that individual performance and fitness-related traits of some fishes are impacted by elevated CO<sub>2</sub>, but there is also the possibility that phenotypic plasticity and/or genetic adaptation could buffer these impacts over the timescales at which CO<sub>2</sub> levels are rising. Therefore, we consider (i) the evidence for acclimation to elevated CO<sub>2</sub> through phenotypic plasticity and (ii) the evidence for genetic variation in phenotypic responses that could provide the raw material for fish populations to adapt to elevated CO<sub>2</sub>. To date, only a handful of studies have attempted to test the adaptive potential of fish populations to elevated CO<sub>2</sub>. We use these as case studies to describe our current understanding of this important topic and to identify major knowledge gaps for future research. Finally, we consider whether there is evidence for artificial selection and adaptation to high CO<sub>2</sub> in aquaculture, which could be informative for predicting the potential for fish populations to adapt to rising CO<sub>2</sub> in the wild.

## 2. EVOLUTIONARY CO<sub>2</sub> HISTORY OF TELEOSTS

The geological timescale over which the fishes evolved commenced over 400 million years ago (Ma: Megaannum or million years ago) during the Silurian Period and the Devonian “Age of the Fishes” (Fig. 1). The earliest

fossils of the cartilaginous fishes date back 430 Ma (Märss and Gagnier, 2006), and the start of the Devonian (419 Ma) marked when jawed fishes divided into three discrete groups including the placoderms (now extinct), early cartilaginous fishes, and bony fishes. This geological time period spanning several hundred million years included four of the Earth's five major mass extinction events: the Late Devonian extinction 375–360 Ma, the Permian Mass Extinction or “The Permian Crisis” 252 Ma, the Triassic-Jurassic Extinction 201.3 Ma, and the Cretaceous-Paleogene (K-Pg) Extinction or the “End Cretaceous” 66 Ma. Over this time period, the atmosphere also experienced dramatic changes in oxygen ( $O_2$ ; declining below 15% at times), global cooling and heating events, and elevated  $CO_2$  levels that, for an extended period of time, exceeded 4000 ppm (Fig. 1). Yet, also during this time, the ray-finned fishes—to which modern teleost fishes belong—diverged from the lobe-finned fishes, and success continued for the following 150 Myr while atmospheric  $O_2$  levels were rising (nearing 30%, compared to today's 21%) and  $CO_2$  levels were falling (from over 4000 ppm to below 500 ppm) (Fig. 1). Then, the Permian Mass Extinction marked a turning point in the evolution of the fishes. The lobe-finned fishes were, for the most part, eliminated from the water (Clack, 2007). Modern lineages moved onto land, leaving ray-finned fishes to expand and exploit the habitats left behind by the lobe-finned fishes (Clack, 2007). Even though the Permian Mass Extinction resulted in a loss of 96% of all marine fish species—of all animals for that matter—this transition signified one of the most profound radiation events among the teleost fishes (Randall et al., 2014). This radiation occurred despite the fact that volcanic activity was causing atmospheric  $CO_2$  levels to rise again while  $O_2$  levels were falling and global temperatures were rising (Fig. 1).

From the start of the cretaceous period 145 Ma, atmospheric  $O_2$  levels rose and  $CO_2$  levels steadily fell. As  $O_2$  levels neared today's 21%, atmospheric  $CO_2$  dropped to around 1000 ppm, and global temperatures were no longer in a warming phase. The K-Pg Extinction event occurred 66 Ma, potentially due to the impact of an asteroid or comet. This mass extinction resulted in a loss of more than 75% of all species on Earth, including the dinosaurs. Despite this, more than 80% of the cartilaginous fishes and more than 90% of teleost fishes survived (Patterson, 1993), even though some fish populations suffered high mortality (Zinsmeister, 1998). After the K-Pg Extinction 66 Ma, atmospheric  $CO_2$  remained below 1000 ppm, falling below 400 ppm by 25 Ma, where it has remained to the current day (Masson-Delmotte et al., 2013). The Early Eocene Climatic Optimum 52–50 Ma was the last time atmospheric  $CO_2$  may have reached 1000 ppm (Masson-Delmotte et al., 2013). Collectively, the dramatically changing atmospheric conditions over more than 400 Myr may have played a profound role in shaping the evolution and diversification of fishes (Fig. 1; Clack, 2007; Ward, 2006).

While the CO<sub>2</sub> content of the atmosphere and the ocean have changed dramatically over the 400 Myr evolutionary history of fishes, the rate of change has not been as fast as the contemporary rise in atmospheric CO<sub>2</sub> concentrations due to anthropogenic CO<sub>2</sub> emissions. On the current emissions trajectory, it is projected that atmospheric CO<sub>2</sub> will exceed 900 ppm by the end of this century (Collins et al., 2013), a level not seen for at least 50 Myr, increasing at a rate at least an order of magnitude faster than at any time in the past 66 Myr (Zeebe et al., 2016).

### 3. ADAPTATION TO HIGH AND VARIABLE CO<sub>2</sub> HABITATS

Among the approximately 34,000 fish species, ranging from primitive species (e.g., lampreys and hagfishes), cartilaginous fishes (approximately 1000 extant named species in the class Chondrichthyes), and modern and more derived teleost fishes (approximately 26,000 named species in the class Actinopterygii, infraclass Teleostei), there are numerous examples of species living in high/variable CO<sub>2</sub> habitats, perhaps making it possible to ascertain their adaptations to these habitats. Indeed, given the variable atmospheric and aquatic CO<sub>2</sub> conditions that fishes have experienced over their evolutionary history (described above), it is not surprising that some species have (pre) adaptations to compensate for low O<sub>2</sub> and/or high CO<sub>2</sub> conditions in some aquatic habitats that would normally preclude fish life. Changes in environmental O<sub>2</sub>, pH, and/or temperature can dramatically affect O<sub>2</sub> uptake, transport, delivery, and CO<sub>2</sub> removal in fishes. There is an intimate interaction between O<sub>2</sub> and CO<sub>2</sub> transport at the gills, and in other tissues, due to their interactions with hemoglobin within the red blood cells (RBCs), which can vary by species (Brauner and Rummel, 2011; Jensen et al., 1993). Generally speaking, most cartilaginous fishes, such as sharks, skates, and rays, can efficiently compensate an acid-base disturbance due to the buffering capacity of their blood and plasma (Berenbrink et al., 2005). Moreover, most sharks, skates, and rays possess relatively pH-insensitive hemoglobins (Berenbrink et al., 2005; Wells et al., 1992), meaning that a pH disturbance associated with an acidosis, such as elevated PCO<sub>2</sub>, may not compromise O<sub>2</sub> transport in the way we understand for teleost fishes. Modern teleost fishes have a different physiological response to an acidosis than the cartilaginous fishes. Teleosts possess extremely pH-sensitive hemoglobins (probably evolving nearly 400 Ma in basal Actinopterygians) and low buffering capacity at the blood and plasma. Adrenergically-activated transporters on the RBCs help to regulate pH, and plasma-accessible carbonic anhydrase in select locations enhances O<sub>2</sub> release from the tissues during an acidotic stress, such as elevated PCO<sub>2</sub> (Berenbrink et al., 2005; Randall et al., 2014; Rummel et al., 2013).

These traits result in an enhanced capacity for O<sub>2</sub> transport, especially during conditions that would normally preclude efficient O<sub>2</sub> uptake. Indeed, these physiological traits may have facilitated the radiation of the teleosts throughout the Triassic and Jurassic periods, despite elevated PCO<sub>2</sub>, low O<sub>2</sub>, and warm temperatures (Fig. 1; Randall et al., 2014).

Ultimately, the adaptations that are in place to maintain O<sub>2</sub> transport and CO<sub>2</sub> removal, even under extreme conditions, may be responsible for the initial and continued (i.e., via modifications in swimming and feeding; Near et al., 2012) success of the fishes (Randall et al., 2014). Many of these adaptations have been investigated in both experimentally and naturally high CO<sub>2</sub> settings, and include over 200 studies to date (Hannan and Rummer, 2018). However, few studies have focused on the mechanisms required for maintaining O<sub>2</sub> transport under an acidosis, such as elevated PCO<sub>2</sub> (Hannan and Rummer, 2018).

Physiological responses to elevated PCO<sub>2</sub> have been investigated under laboratory conditions in an array of taxa, including primitive fishes, both freshwater and marine, and even air-breathing fishes (Chapter 2, Vol 37: Tresguerres et al., 2019; Chapter 3, Vol 37: Brauner et al., 2019; Chapter 8, Vol 37: Skov, 2019). For example, hagfish have been well studied to understand the evolution of O<sub>2</sub> and CO<sub>2</sub> transport and the relationship to acid-base balance and ion/water balance, as the hagfish retains several key ancestral traits associated with the most primitive vertebrates (Baker et al., 2015). In fact, it was the hagfish's superior capacity to tolerate and compensate for acidosis associated with elevated PCO<sub>2</sub> that gave concrete evidence that the gills first evolved to serve acid-base balance and then later for O<sub>2</sub> transport (Baker et al., 2015). While hagfish are known to create high CO<sub>2</sub> microhabitats due to their lifestyle and feeding habits that involve burrowing into sediment and/or rotting animal flesh, other fish species have been investigated due to their more obvious tendencies to inhabit elevated CO<sub>2</sub> environments. Highly productive lagoons and tide-pools as well as freshwater ponds and rivers (e.g., the Amazon) can diurnally and seasonally experience PCO<sub>2</sub> levels that are nearly 80-times what is predicted in the atmosphere by the end of the 21st century from anthropogenic CO<sub>2</sub> emissions (Ultsch, 1996). To tolerate and compensate for these conditions, species like swamp eel (*Synbranchus marmoratus*), armored catfish (*Pterygoplichthys pardalis*), and white sturgeon (*Acipenser transmontanus*) can preferentially regulate intracellular pH without added metabolic costs (Baker and Brauner, 2012; Brauner et al., 2004; Heisler, 1982). This strategy may be what allows such species that regularly encounter very high environmental PCO<sub>2</sub> to survive and thrive in such habitats. Some of the facultative and obligate air breathing fishes (e.g., *Arapaima gigas* and *Lipossarcus pardalis*) that inhabit diurnally and seasonally extreme habitats (e.g., the Amazon) have also been well studied, not only because of the elevated PCO<sub>2</sub> and often low O<sub>2</sub> of their environment, but also because of

the division of tasks between the gill and the air breathing organ (e.g., swimbladder) for gas exchange and acid-base balance (Brauner and Val, 1996). These species use the characteristics of their hemoglobin to ensure O<sub>2</sub> uptake is not compromised for the sake of CO<sub>2</sub> removal (Brauner and Val, 1996), which would normally happen in other teleost fishes, because hemoglobin-O<sub>2</sub> binding requires the proton resulting from bicarbonate dehydration to form CO<sub>2</sub> for release via the gill, thus tightly coupling O<sub>2</sub> and CO<sub>2</sub> exchange (Brauner and Rummer, 2011). More derived species may either move to avoid daily or seasonal periods of high environmental PCO<sub>2</sub> or tolerate these episodic events through morphological adjustments (e.g. gill remodeling) and physiological compensation via exchanging bicarbonate from the environment for a counter-ion to correct extracellular pH (Deigweiher et al., 2008; Heuer and Grosell, 2014). But bicarbonate-mediated compensation is neither sustainable over the long term, nor under extreme hypercapnic conditions, as extracellular pH compensation is limited by the amount of bicarbonate that can be exchanged (Brauner and Baker, 2009; Heisler, 1986). Consequently, there are limits to the duration of exposure to extremely high CO<sub>2</sub> that teleosts can both tolerate and function effectively. Diverse teleost communities live around natural CO<sub>2</sub> seeps (e.g. Munday et al., 2014) in locations where PCO<sub>2</sub> is similar to that predicted for the end of the century (i.e., ~1000  $\mu$ atm). However, fewer species occur near intense vents (P. L. Munday, personal observation) where PCO<sub>2</sub> ranges between 5000 and 10,000  $\mu$ atm (Fabricius et al., 2011) and many that do enter these areas are mobile feeders on seagrass beds that dominate these extreme CO<sub>2</sub> locations. Therefore, they likely experience very high CO<sub>2</sub> for short periods of time but most species are not permanently exposed to extreme CO<sub>2</sub> levels.

Beyond the mechanisms that more derived teleost fishes may use to compensate a pH disturbance under elevated CO<sub>2</sub> conditions, a mechanism by which O<sub>2</sub> delivery is maintained, if not enhanced, under such scenarios has been described, including in what would be considered a CO<sub>2</sub> sensitive species (i.e., rainbow trout; Alderman et al., 2016; Rummer and Brauner, 2011; Rummer et al., 2013). With pH sensitive hemoglobins, a reduction in pH not only reduces the affinity of hemoglobin for O<sub>2</sub> but also the carrying capacity (combined Bohr-Root effects). Under elevated CO<sub>2</sub> conditions, the ensuing acidosis often initiates an adrenergic response whereby catecholamines (e.g., adrenaline, noradrenaline) are released into circulation, and among other functions, also serve to activate ion exchangers (e.g., sodium/proton Na<sup>+</sup>/H<sup>+</sup> exchange;  $\beta$ NHE) on the RBCs to protect intracellular pH and therefore O<sub>2</sub> transport (Rummer et al., 2013). Maintaining RBC pH increases hemoglobin-O<sub>2</sub> affinity and therefore safeguards O<sub>2</sub> uptake at the gill, despite an acidosis. However, at select locations (e.g., red muscle and heart), plasma-accessible carbonic anhydrase short-circuits this protective

mechanism. The RBC is, therefore, reacidified, hemoglobin-O<sub>2</sub> affinity decreases, and O<sub>2</sub> is released to the tissue at those select locations (Rummer *et al.*, 2013). In the species that have been investigated (i.e., salmonids) this mechanism may also be operating under a mild acidosis, not necessarily requiring catecholamine release to trigger the NHE. Moreover, this may be the mechanism by which some teleosts can maintain or even enhance aerobic metabolic scope under mildly elevated CO<sub>2</sub> conditions (Hannan and Rummer, 2018). However, this does not apply to cartilaginous fishes. As mentioned above, they have a relatively pH-insensitive hemoglobin and no adrenergically stimulated Na<sup>+</sup>/H<sup>+</sup> transporters on their RBCs, precluding this mechanism from benefitting O<sub>2</sub> delivery. Despite the ability of most fishes to maintain O<sub>2</sub> delivery under elevated PCO<sub>2</sub> (reviewed in Hannan and Rummer, 2018; Lefevre, 2016; Chapter 6, Vol 37: Lefevre, 2019), effects on growth, survival and swimming performance have been observed in a range of species, especially in early life stages (Cattano *et al.*, 2018; Heuer and Grosell, 2014; Chapter 9, Vol 37: Munday *et al.*, 2019a), indicating that processes other than O<sub>2</sub> delivery become limiting at higher CO<sub>2</sub> levels. Regardless of the exact mechanism, understanding the link between the physiological performance of a diverse range of extant fishes in elevated CO<sub>2</sub> and their evolutionary history may help in predicting which species will respond to higher CO<sub>2</sub> conditions in the future.

#### **4. PHENOTYPIC PLASTICITY AND GENETIC ADAPTATION TO FUTURE CO<sub>2</sub> LEVELS**

Experiments have revealed that some fish species are evidently sensitive to predicted CO<sub>2</sub> levels over the next few hundred years (Cattano *et al.*, 2018). These empirical assessments of CO<sub>2</sub> sensitivity are aptly viewed as an important first step, because they can distinguish CO<sub>2</sub>-tolerant from potentially vulnerable species, affected from unaffected traits, and identify the most sensitive ontogenetic stages (Baumann, 2019). But a species' vulnerability to predicted future CO<sub>2</sub> levels is not just determined by its contemporary CO<sub>2</sub> sensitivity. Apart from indirect effects (i.e., altered food web or species interactions), the most important question is whether phenotypic plasticity and/or evolution can buffer the direct effects of elevated CO<sub>2</sub> over the timescale at which the environment is changing (Munday *et al.*, 2013b; Reusch, 2014; Sunday *et al.*, 2014). Therefore, the overall vulnerability to future elevated CO<sub>2</sub> levels might be thought of as the sensitivity to elevated CO<sub>2</sub> identified in experiments, less the adaptive potential from phenotypic plasticity and genetic adaptation.

$$\text{CO}_2 \text{ vulnerability} = \text{CO}_2 \text{ sensitivity} - (\text{phenotypic plasticity} + \text{genetic adaptation}) + \text{indirect effects}$$

#### 4.1. Within and Between Generation Plasticity to High CO<sub>2</sub>

While exposure to elevated CO<sub>2</sub> may have negative effects in the short-term, individuals may be able to adjust to these conditions over the longer term as a result of plasticity in physiological processes that underpin growth, survival, reproduction, and behavior. In general, there is limited evidence for within-generation acclimation of fish to elevated CO<sub>2</sub>. Instead, most experimental studies that conduct sampling at varying time points tend to find that the species or trait of interest is either affected or unaffected by elevated CO<sub>2</sub> and remains that way through time (e.g. [Davis et al., 2018](#); [Munday et al., 2013a](#); [Sundin et al., 2017](#)). In other words, most studies report either tolerance or sensitivity to elevated CO<sub>2</sub> ([Hamilton et al., 2017](#); [Schmidt et al., 2017](#)), not an initial high sensitivity that is then buffered by phenotypic plasticity. In one of the few clear exceptions, [Lopes et al. \(2016\)](#) found that exposure to elevated CO<sub>2</sub> for 7 days affected group cohesion in larval sand smelt (*Atherina presbyter*), but it was no longer affected after 21 days of exposure. One reason for the paucity of studies showing within-generation plasticity to elevated CO<sub>2</sub> is that few studies have compared the same traits over short vs long exposure durations. It is possible that more examples of within-generation plasticity and acclimation to elevated CO<sub>2</sub> will emerge if greater attention is given to repeat sampling of CO<sub>2</sub>-sensitive traits over relevant timeframes. Indeed, in the only study of its type to date, [Schunter et al. \(2018\)](#) found 184 genes were differentially expressed in spiny damselfish that had been exposed to elevated CO<sub>2</sub> for 4 days vs 5 month before sampling. The differentially expressed genes in the fish exposed to elevated CO<sub>2</sub> for 4 days were significantly enriched in ATPase-related processes, indicating an initial physiological cost to high CO<sub>2</sub> exposure, which does not persist over the longer term.

In contrast to the limited evidence for within-generation plasticity and acclimation to elevated CO<sub>2</sub>, there is ample evidence that exposure to elevated CO<sub>2</sub> in one generation can affect the performance of future generations. [Miller et al. \(2012\)](#) were the first to show that parental effects can improve the performance of fish early life stages under elevated CO<sub>2</sub>. Juvenile cinnamon anemonefish (*Amphiprion melanopus*) had a higher metabolic rate, slower growth rate, and higher mortality when reared in elevated CO<sub>2</sub> compared to controls. However, these negative effects of elevated CO<sub>2</sub> were completely ameliorated in offspring whose parents had been exposed to elevated CO<sub>2</sub> for several months prior to the start of the breeding season. This result was recently repeated by [Jarrold and Munday \(2019\)](#), who showed that beneficial parental effects on offspring growth and survival were of similar magnitude to the beneficial effects arising from diel cycles in the CO<sub>2</sub> treatment (Chapter 9, Vol 37: [Munday et al., 2019a](#)). More efficient acid-base regulatory processes and/or mitochondrial function when both parents and offspring experienced

the same high CO<sub>2</sub> conditions could potentially explain the restoration of juvenile resting metabolic rate to control levels and the corresponding improvement in growth and survival of juveniles from high CO<sub>2</sub> parents (Miller et al., 2012). Murray et al. (2014) found similar beneficial parental effects in a wild population of Atlantic silversides (*Menidia menidia*). Offspring produced from wild spawners repeatedly throughout the spawning season became progressively more tolerant to elevated CO<sub>2</sub>, which coincided with the natural pH decline (CO<sub>2</sub> increase) in the adult's habitat during the spawning season. Here again, matching environmental conditions experienced by parents and offspring appeared to improve performance of the offspring in high CO<sub>2</sub>.

Baumann et al. (2018) took the Atlantic silverside story a step further. By combining results of experiments conducted at various times throughout the season they showed that juvenile tolerance to high CO<sub>2</sub> first climbs in conjunction with rising CO<sub>2</sub> in the parental spawning habitat, but then declines again late in the season, most likely due to reduced maternal egg provisioning and thus reduced egg quality late in the season. This additional knowledge is important, because it illustrates that parental effects are not a "magic bullet" and can interact with energetic constraints and other environmental factors. Indeed, in an experimental demonstration of this issue, Stiasny et al. (2018) showed that the presence of beneficial parental effects may depend on the energy supply available to the offspring. Reduced survival of larval Atlantic cod (*Gadus morhua*) in elevated CO<sub>2</sub> was partly compensated by parental exposure to high CO<sub>2</sub> exposure, but only when ample food was available for the larvae. This interaction with food supply suggests that surplus metabolic resources may be needed to elicit beneficial transgenerational effects. In contrast to these positive parental effects, parental exposure to elevated CO<sub>2</sub> had a negative effect on the survival and growth of juvenile threespine stickleback (*Gasterosteus aculeatus*) (Schade et al., 2014), demonstrating that not all parental effects are beneficial, and that stressful conditions in one generation can have negative carryover effects to future generations. Overall, these studies highlight the importance of considering parental and transgenerational effects when attempting to determine how fish early life stages will respond to rising CO<sub>2</sub> levels.

Only a handful of studies have tested for beneficial parental effects on behavior in elevated CO<sub>2</sub>. Welch et al. (2014) found no evidence for parental effects in behavioral responses to elevated CO<sub>2</sub> in juvenile spiny damselfish. Juveniles exhibited impaired responses to alarm cues and altered behavioral lateralization, irrespective of whether their parents had been exposed to high CO<sub>2</sub> or not. Similar results were observed in clownfish, where elevated CO<sub>2</sub> impaired antipredator behavior of juveniles, regardless of the CO<sub>2</sub> treatment of their parents (McMahon et al., 2018). By contrast, Allan et al. (2014) found

that parental exposure to elevated CO<sub>2</sub> reduced the negative effects of high CO<sub>2</sub> on the escape performance of juvenile fish, especially the proportion of fish that responded to the threat stimulus, indicating that beneficial parental effects may occur in some traits and in some contexts, but not others.

The parental effects described above could occur through a number of different mechanisms. Parents can influence the phenotype of their offspring by differences in nutritional provisions of eggs (Snyder et al., 2018), the transmission of hormones, proteins and other somatic factors, and potentially by the transfer of epigenetic marks, including DNA methylation and histones (Munday, 2014). As yet, the mechanisms by which parental exposure to elevated CO<sub>2</sub> influences the performance of offspring under high CO<sub>2</sub> have not been identified, although new molecular approaches could help answer this question. For example, Schunter et al. (2018) found that patterns of gene expression altered by within-generation exposure to elevated CO<sub>2</sub> returned to baseline levels following parental exposure to elevated CO<sub>2</sub>. This suggests a fundamental resetting of developmental pathways when offspring experience the same high CO<sub>2</sub> condition as their parents. Epigenetic mechanisms could be involved, and this will be a fruitful area for further research.

#### 4.2. Heritability and Selection to High CO<sub>2</sub> in Fishes

It is possible that fishes could adapt to future higher CO<sub>2</sub> levels. To illustrate this point, consider an experiment where exposure to elevated CO<sub>2</sub> reduced the survival of an experimental population by half. This doubling in mortality would be concerning to an ecologist, but an evolutionary biologist would point out that 50% of the population appeared to be CO<sub>2</sub> tolerant and would therefore have a fitness advantage in future environments. Thus we might observe a gradual increase in the frequency of the alleles conferring CO<sub>2</sub> tolerance, registering as an adaptive evolutionary response of the population. In theory, CO<sub>2</sub> resistant genotypes could also arise randomly from new mutations (*de novo*), but this pathway is likely limited to single cell organisms with generation times of days to weeks (Lohbeck et al., 2012; Reusch, 2014). For fish and other metazoans with generations of months to decades, it is the genetic variation that currently exists in natural populations that may or may not allow them to adapt to high CO<sub>2</sub> (Sunday et al., 2014). Generation length and standing genetic diversity are important parameters, but to predict the potential, strength and direction of evolutionary responses to high CO<sub>2</sub> oceans, we still need to know more.

To approach the task, we can start with the *breeders equation* (Falconer and Mackay, 1996).

$$\Delta Z_i = h^2_i \times S_i$$

This equation calculates the change in trait  $i$  from one generation to the next ( $\Delta Z$ ) as the traits heritability ( $h^2$ ) multiplied by the strength of selection (selection differential,  $S$ ). Heritability measures the proportion of phenotypic variation in a trait that is attributed to additive genetic variance; hence, strongly heritable traits can potentially evolve quickly under selection, whereas nonheritable traits will not respond. Fortunately, estimating the heritability part of the breeders equation is relatively straightforward. It has routinely been done using cross-breeding designs, parent-offspring regressions or other approaches as long as they allowed tracking genetic relatedness in experimental populations (Mousseau and Roff, 1987). Unfortunately, published trait heritabilities specific to high CO<sub>2</sub> environments are still rare (Sunday et al., 2014) and remain scarcer still for fish.

One pioneering experiment subjected newly fertilized Atlantic silverside embryos from a large number of wild spawners to high CO<sub>2</sub> (~2300  $\mu\text{atm}$ ) and then quantitatively sampled all perished and surviving individuals for 15 days posthatch (Malvezzi et al., 2015). Afterward, all parents and offspring were genotyped using microsatellite markers, which allowed reconstructing parent–offspring pedigrees and thereby revealed that genetically related individuals were more similar in their survival than unrelated ones. Because the additive genetic variance comprised a fifth of the total phenotypic variance, the heritability of survival at high CO<sub>2</sub> was estimated to be 0.20. This value is remarkable only for being utterly common in complex, polygenic traits like survival, growth or fecundity across taxa and irrespective of the agents of selection (Mousseau and Roff, 1987).

Similar heritabilities of larval mortality and growth under contrasting CO<sub>2</sub> conditions were recently estimated for California grunion (*Leuresthes tenuis*) (Tasoff and Johnson, 2019). Here, the authors used a classic cross-breeding design to produce 31 offspring families with different degrees of relatedness, which were reared for 14 days post hatch under control (650  $\mu\text{atm}$ ) vs high (1630  $\mu\text{atm}$ ) CO<sub>2</sub> conditions. As in the first experiment, a multivariate animal model (Johnson et al., 2010; Kruuk, 2004; Wilson et al., 2007) allowed partitioning phenotypic variances of larval mortality and growth, thereby estimating heritability and maternal effects in these traits. Again, heritabilities for larval mortality were small but significant (0.12–0.18) and even lower for larval growth (0.09–0.10). As is commonly observed in marine species (Marshall et al., 2008), nongenetic maternal effects explained a considerable proportion (26–36%) of the variability in larval mortality.

Apart from mortality and growth, physiological and behavioral traits could affect fitness, too, and therefore respond to selection. In tropical reef fish, high CO<sub>2</sub> exposure is known to elicit abnormal behavior (e.g., loss of predator avoidance or orientation) in many but not all individuals of an experimental population. The fitness relevance of these effects was demonstrated in

a combined laboratory and field study, showing that CO<sub>2</sub> affected damselfish juveniles (*Pomacentrus wardi*) suffered higher predation mortalities when transplanted to the field (44%) than their CO<sub>2</sub> tolerant conspecifics (32%) (Munday et al., 2012). The heritability of predator avoidance behavior (response to chemical alarm cues) was subsequently estimated in a different damselfish model, the spiny chromis (*Acanthochromis polyacanthus*) (Welch and Munday, 2017). Here, the additive genetic component of the phenotypic trait variance was determined via father-offspring regressions, an alternative to unworkable cross-breeding designs in species with parental offspring care. Intriguingly, avoidance behavior had a high heritability ( $h^2 = 0.56\text{--}0.65$ ) in both wild and laboratory reared populations only when estimated after acute exposures (4 days) to elevated CO<sub>2</sub> (750  $\mu\text{atm}$ ). After prolonged high CO<sub>2</sub> exposure (42 days), however, behavioral responses appeared non-heritable ( $h^2 = -0.01\text{--}0.02$ ). This result came about because variation in phenotypic responses to elevated CO<sub>2</sub> that were present after 4 days exposure to elevated CO<sub>2</sub> were absent after 42 days. Thus, it seems that nonadaptive plasticity acted to obscure genetic variance in the trait and thereby reduce its heritability. The reality of continuous high CO<sub>2</sub> exposure in the future may be that reduced phenotypic variation could constrain the evolutionary potential in fitness-relevant traits.

Estimating trait heritabilities under high CO<sub>2</sub> conditions is a required first step toward anticipating evolutionary responses. Appropriate experimental approaches exist (Munday et al., 2013b; Sunday et al., 2014), but have been applied in just a handful of cases. So far, the sparse empirical evidence for fish suggests that fitness relevant survival and growth responses to high CO<sub>2</sub> conditions have a small but significant genetic basis on which selection could act. It is the other half of the breeders equation—estimating selection—where the knowledge gap is widest and available tools are most wanting.

#### 4.3. Selection Differentials

Almost by definition, today's laboratory experiments determine species sensitivities to average CO<sub>2</sub> levels as they may occur decades to centuries from now. Where responses occur, selection differentials could be calculated (Malvezzi et al., 2015; Munday et al., 2012), but these could still be unrealistic because average CO<sub>2</sub> levels are rising in a gradual process spanning dozens of generations. Consider that current models predict a maximum rise in atmospheric CO<sub>2</sub> levels to  $\sim 2000\mu\text{atm}$  by the year 2300 (Caldeira and Wickett, 2003), i.e., a  $\Delta\text{CO}_2$  of 1600  $\mu\text{atm}$  in 280 years. For a hypothetical fish species with a generation span of 3 years this amounts to a 17  $\mu\text{atm}$  increase in average CO<sub>2</sub> level from one generation to the next. Yet, individuals within each generation will also be exposed to PCO<sub>2</sub> variation on daily, seasonal and periodic

(e.g. upwelling) timescales that will be far greater than the year-by-year increase in average  $\text{PCO}_2$ . It is possible that the gradual nature of the rise in average  $\text{CO}_2$  levels may thus impose very small selection differentials, and in combination with small trait heritabilities, result in little to no evolutionary change for many decades, until the average  $\text{PCO}_2$  or the  $\text{PCO}_2$  peaks associated with natural variation rise to levels that have a meaningful effect on fish fitness. Moreover, beneficial phenotypic plasticity (as discussed in [Section 4.1](#)) will conceivably weaken the impetus for genetic adaptation further. The concern here is that eventually environmental changes will exceed the acclimation threshold, which could trigger unexpected and severe productivity losses in the short term, until adaptation finally catches up over the long term ([Bell, 2013](#)).

The generation times of fish and the incremental nature of rising  $\text{CO}_2$  levels render real-time experiments largely useless for quantifying realistic selection. Computer models could be an alternative and if appropriately parameterized offer scenarios for how specific combinations of trait heritabilities  $\times$  selection differentials may play out over dozens to hundreds of fish generations. Some insights may also come from examining intraspecific local adaptations in species distributed across spatial  $\text{CO}_2$  gradients ([Calosi et al., 2017](#); [Sunday et al., 2011](#)). Such gradients are abundant in nature and exist with depth or between regions that differ in upwelling intensity ([Kelly et al., 2013](#); [Vargas et al., 2017](#)), at natural  $\text{CO}_2$  seeps ([Fabricius et al., 2011](#)) or across the general nearshore–offshore axis of diminishing natural  $\text{CO}_2$  variability ([Baumann, 2019](#)). The value of these so-called “space-for-time” approaches is a long-standing topic of scientific debate, as these approaches may elucidate whether evolution can occur and what adapted traits look like, while leaving the question how long this adaptive process takes unanswered ([Baumann and Conover, 2011](#)). Furthermore, the small spatial extent of some  $\text{CO}_2$  gradients, such as  $\text{CO}_2$  seeps, means that the very populations under selection are also constantly being swamped with genotypes from the surrounding low  $\text{CO}_2$  environment, thus overriding the potential for genetic change ([Munday et al., 2013b](#)).

#### 4.4. Genetic Correlations

Estimating the evolutionary potential of a single trait to a single environmental stressor is challenging enough, but realistic predictions for multiple traits and stressors are still more complicated. For one, fish will not adapt to high  $\text{CO}_2$  in isolation, but to future high  $\text{CO}_2$  environments that will also be warmer, less oxygenated, and likely exhibit altered food webs and habitats ([Enochs et al., 2015](#), Chapter 9, Vol 37: [Munday et al., 2019a](#)). For example, some tropical reef fish may be capable of acclimating and adapting to warming ([Donelson et al., 2012](#); [Munday et al., 2017](#)), but could still fail to

overcome rapid climate degradation of their coral reef habitats ([Jones et al., 2004](#); [Pratchett et al., 2008](#)). For multiple stressors acting on a single trait (e.g., survival), exploring underlying genetic correlations will be crucial. Genetic variance among individuals in a population could vary such that CO<sub>2</sub> tolerant genotypes also survive better at higher temperatures (positive genetic correlation), thus leading to an accelerated adaptive response. However, negative genetic correlations are equally plausible, which would dampen or impede adaptation in a multistressor environment ([Munday et al., 2013b](#); [Sunday et al., 2014](#)).

Related but separate from genetic correlations are evolutionary trade-offs between multiple traits. Darwinian fitness measures an individual's contribution of progeny to the next generation, and this overall outcome is the product of many fitness-related traits. These traits are likely not free to evolve independently, but involve trade-offs that lead to evolutionary constraints. A textbook analogue from fish life history is the strong negative coupling between annual mortality rate and maturation age, where combinations of (i) late maturation/high mortality or (ii) early maturation/low mortality simply cannot arise due to evolutionary trade-offs in reproduction ([Charnov and Berrigan, 1991](#)). A new study by [Laubenstein et al. \(2019\)](#) has demonstrated such a trade-off between fitness-related traits in reef fishes under the combined climate change conditions of elevated CO<sub>2</sub> and warming. In the spiny damselfish there was a negative correlation between physiological performance (aerobic scope) and behavioral performance (appropriate response to a predation-risk cue), but only in fish reared in the combined elevated CO<sub>2</sub> and high temperature treatment. This negative correlation means that individuals with high aerobic scope (a measure of aerobic performance) exhibit the most impaired response to a predation-risk cue, and visa-versa. There is a trade-off because individuals can have high performance in one trait, but not both. If this trade-off has a corresponding genetic correlation, it would be expected to constrain the evolution of one trait or both under future climate change conditions.

## 5. SELECTION IN AQUACULTURE

Before the term “ocean acidification” was even coined, aquaculturists had already been studying fish sensitivities to high CO<sub>2</sub> for decades ([Ishimatsu et al., 2008](#)), recognizing that extreme CO<sub>2</sub> environments (>10,000  $\mu\text{atm}$ ) are a frequent metabolic byproduct of maintaining fish in high densities in confined enclosures ([Ellis et al., 2017](#), Chapter 8, Vol 37: [Skov, 2019](#)). In fact, the apparent CO<sub>2</sub> tolerance of fish juveniles and adults likely contributed to an early—but now outdated—notion that rising CO<sub>2</sub> levels from anthropogenic emissions will not affect fish. It is now recognized that

aquaculture conditions provide a unique opportunity to better understand the effects of high CO<sub>2</sub> on fishes (Ellis et al., 2017), and few areas are as ripe for discovery as the physiological and genomic underpinnings of acclimation and adaptation to high CO<sub>2</sub> environments. In farmed species with closed life cycles artificial selection to high CO<sub>2</sub> is likely, thereby presenting opportunities to quantify how selection decreases genetic diversity and how trait and genetic correlations may accelerate or impede adaptive change. For example, a recent study examined quantitative trait loci (QTL) in the genome of wild, farmed, and wild × farmed Atlantic salmon (*Salmo salar*), finding that six QTLs explained 10–15% of the phenotypic variation in length, weight, and condition factor, while a seventh QTL was significantly related to survival (Besnier et al., 2015). Interestingly, within one of the QTL regions of the genome the analyses revealed strong allelic correlations (linked genomic variants) most likely caused by strong artificial selection “sweeping” (removing) large parts of the genetic diversity in farmed animals. For salmon, this is a recognized conservation concern, given the well-documented accidental interbreeding between farmed and wild strains. There is good evidence for increased CO<sub>2</sub> tolerance in farmed populations of Sydney rock oysters compared with their wild counterparts (Parker et al., 2011), but whether there is similar selection for high CO<sub>2</sub> tolerance in farmed fishes is currently unknown.

Domestication and loss of genetic diversity in aquaculture populations also has profound yet often unacknowledged implications for experimental research on the effects of high CO<sub>2</sub> in general. For practical reasons, many experimental research groups obtain specimens (particularly embryos and larvae) from commercial and laboratory brood stocks to conduct sensitivity assessments of these organisms to climate stressors (e.g. Baumann et al., 2012; Bromhead et al., 2015; Chambers et al., 2014; Stiasny et al., 2018). While these studies are valuable, they are likely not accurate representations of the genetic diversity and thus evolutionary potential of species and populations as they currently exist in the wild. The widespread potential of marine fish to precondition offspring to parental environments (i.e., Section 4.1 above) compounds the problem, because brood stocks are not only genetically altered, but also often experience constant temperature, CO<sub>2</sub>, light, and feeding conditions that do not resemble conditions in the wild (Baumann, 2019). This additional uncertainty needs to be more clearly communicated and accounted for in an attempt to assess the sensitivity of fish to future CO<sub>2</sub> levels, and their adaptive potential.

## 6. KNOWLEDGE GAPS AND FUTURE DIRECTIONS

The evolutionary history of fishes spans periods of geological time where atmospheric CO<sub>2</sub> was much higher than today, and some fishes continue to

occupy habitats that can experience very high PCO<sub>2</sub> (>10,000  $\mu$ atm). By contrast, other extant fishes appear to be adapted to the low atmospheric CO<sub>2</sub> conditions (<400  $\mu$ atm) that have persisted for the last 25 Ma. The physiological processes that make some species sensitive and others tolerant to projected future CO<sub>2</sub> levels (~1000–2000  $\mu$ atm) are still not well understood. Phylogenetically controlled comparative studies could be especially useful in establishing what processes and mechanisms are responsible for these differences in CO<sub>2</sub> tolerance among fishes and how they are related to the evolutionary history of modern fishes.

The greatest knowledge gap in predicting the effects of future higher CO<sub>2</sub> levels on marine and freshwater fishes is the limited understanding of their adaptive potential to the rapid rise in PCO<sub>2</sub> that will occur over coming decades. An increasing number of studies are testing for adaptive plasticity to higher CO<sub>2</sub> levels in fish populations, both within and between generations. These studies illustrate the importance of considering parental effects, because negative effects of elevated CO<sub>2</sub> observed in offspring may be mitigated when parents experience the same environmental conditions. Nevertheless, parental effects must not be viewed as a “magic bullet”, because not all traits exhibit improved performance, and there are also examples of negative carry-over effects from parental exposure to high CO<sub>2</sub>. One issue in studying parental and transgenerational effects is that most experimental designs used to date cannot adequately separate apparent transgenerational effects from developmental plasticity (Donelson et al., 2018). More complex experimental designs are needed if future studies are to disentangle true transgenerational effects from developmental plasticity induced by exposure of critical early life stages (e.g. eggs and embryos) to high CO<sub>2</sub> (Donelson et al., 2018). We also have limited understanding of the physiological mechanisms underpinning most examples of parental and transgenerational effects, although modern molecular methods are being used to address this issues (Ryu et al., 2018; Schunter et al., 2018) and could provide unprecedented insight into the mechanisms of within and between generation plasticity.

Despite a rich theory and well-developed experimental methods in quantitative genetics (e.g. Lynch and Walsh, 1998), only a handful of studies have tested for genetic variation in CO<sub>2</sub> sensitive traits, which might enable fish to adapt to projected future CO<sub>2</sub> levels. This is a serious knowledge gap that needs a concerted research effort to overcome. Without basic information on genetic variation in fitness-associated traits and the strength of selection, it is not possible to make informed decisions about the impacts of elevated CO<sub>2</sub> on fish populations over the timeframes that CO<sub>2</sub> is changing. Most fish populations will experience dozens of generations before PCO<sub>2</sub> reaches levels that cause negative populations effects. Moreover, higher CO<sub>2</sub> levels will be accompanied by higher temperatures, and even less is known about genetic

correlations of traits in response to different environmental stressors that could either constrain or facilitate the pace of adaptive responses.

Meaningful estimates of some important evolutionary parameters, such as selection gradients to intergenerational increases in  $\text{PCO}_2$  will be difficult to measure. Sensitivity analyses will be useful to modeling adaptive responses under a range of selection scenarios. So-called evolutionary rescue models (see Bell, 2013 and associated papers) are likely to be especially powerful in assessing population level responses to high  $\text{CO}_2$  and warming because they can explicitly incorporate both the negative demographic effects of these stressors on the population, and the population genetic response to selection, to predict population trajectories under a range of environmental conditions and evolutionary parameters. While this approach has been used in terrestrial ecosystems, to our knowledge, it has not yet been used to model the trajectory of fish populations in response to global climate change.

Finally, it is simply impossible to assess the evolutionary potential in every fish species, or even a tiny percentage of all species. Therefore, to gain traction it will be important to focus on a few species that are amenable to experimental manipulation in an evolutionary context. Species such as the spiny damselfish, clownfish, Atlantic silverside and salmon that have already been a mainstay of plasticity and adaption research, and for which good molecular resources have been developed, are good candidates. The yellowtail kingfish is emerging as a representative for large pelagic fishes (Munday et al., 2019b). Representative species from polar, deep sea and rocky temperate habitats are needed. Model species have proven enormously successful in testing evolutionary questions in terrestrial biology and are likely to be similarly successful in aquatic ecosystems.

## ACKNOWLEDGMENTS

We thank Erin Walsh for preparing the figures and Michael Jarrold for assistance with the manuscript.

## REFERENCES

Alderman, S.L., Harter, T.S., Wilson, J.M., Supuran, C.T., Farrell, A.P., Brauner, C.J., 2016. Evidence for a plasma-accessible carbonic anhydrase in the lumen of salmon heart that may enhance oxygen delivery to the myocardium. *J. Exp. Biol.* 219, 719–724. <https://doi.org/10.1242/jeb.130443>.

Allan, B.J.M., Miller, G.M., McCormick, M.I., Domenici, P., Munday, P.L., 2014. Parental effects improve escape performance of juvenile reef fish in a high- $\text{CO}_2$  world. *Proc. R. Soc. B Biol. Sci.* 281, 20132179. <https://doi.org/10.1098/rspb.2013.2179>.

Baker, D.W., Brauner, C.J., 2012. Metabolic changes associated with acid-base regulation during hypercarbia in the  $\text{CO}_2$ -tolerant chondrostean, white sturgeon (*Acipenser transmontanus*). *Comp. Biochem. Physiol. A Mol. Integr. Physiol.* 161, 61–68. <https://doi.org/10.1016/j.cbpa.2011.09.002>.

Baker, D.W., Sardella, B., Rummer, J.L., Sackville, M., Brauner, C.J., 2015. Hagfish: champions of CO<sub>2</sub> tolerance question the origins of vertebrate gill function. *Sci. Rep.* 5, 11182. <https://doi.org/10.1038/srep11182>.

Baumann, H., 2019. Experimental assessments of marine species sensitivities to ocean acidification and co-stressors: how far have we come? *Can. J. Zool.* 97, 399–408. <https://doi.org/10.1139/cjz-2018-0198>.

Baumann, H., Conover, D.O., 2011. Adaptation to climate change: contrasting patterns of thermal-reaction-norm evolution in Pacific versus Atlantic silversides. *Proc. R. Soc. B Biol. Sci.* 278, 2265–2273. <https://doi.org/10.1098/rspb.2010.2479>.

Baumann, H., Smith, E.M., 2018. Quantifying metabolically driven pH and oxygen fluctuations in US nearshore habitats at diel to interannual time scales. *Estuaries Coast* 41, 1102–1117. <https://doi.org/10.1007/s12237-017-0321-3>.

Baumann, H., Talmage, S.C., Gobler, C.J., 2012. Reduced early life growth and survival in a fish in direct response to increased carbon dioxide. *Nat. Clim. Chang.* 2, 38–41. <https://doi.org/10.1038/nclimate1291>.

Baumann, H., Cross, E.L., Murray, C.S., 2018. Robust quantification of fish early life CO<sub>2</sub> sensitivities via serial experimentation. *Biol. Lett.* 14, 20180408. <https://doi.org/10.1098/rsbl.2018.0408>.

Bell, G., 2013. Evolutionary rescue and the limits of adaptation. *Philos. Trans. R. Soc. Lond. B Biol. Sci.* 368, 20120080. <https://doi.org/10.1098/rstb.2012.0080>.

Berenbrink, M., Koldkjaer, P., Kepp, O., Cossins, A.R., 2005. Evolution and oxygen secretion in fishes and the emergence of a complex physiological system. *Science* 307, 1752–1757. <https://doi.org/10.1126/science.1107793>.

Besnier, F., Glover, K.A., Lien, S., Kent, M., Hansen, M.M., Shen, X., Skaala, Ø., 2015. Identification of quantitative genetic components of fitness variation in farmed, hybrid and native salmon in the wild. *Heredity* 115, 47–55. <https://doi.org/10.1038/hdy.2015.15>.

Bonduriansky, R., Crean, A.J., Day, T., 2012. The implications of nongenetic inheritance for evolution in changing environments. *Evol. Appl.* 5, 192–201. <https://doi.org/10.1111/j.1752-4571.2011.00213.x>.

Borges, A.V., Abril, G., Darchambeau, F., Teodoro, C.R., Deborde, J., Vidal, L.O., Lambert, T., Bouillon, S., 2015. Divergent biophysical controls of aquatic CO<sub>2</sub> and CH<sub>4</sub> in the World's two largest rivers. *Sci. Rep.* 5, 15614. <https://doi.org/10.1038/srep15614>.

Brauner, C.J., Baker, D.W., 2009. Patterns of acid-base regulation during exposure to hypercarbia in fishes. In: Glass, M., Wood, S.C. (Eds.), *Cardio-Respiratory Control in Vertebrates: Comparative and Evolutionary Aspects*. Springer, Berlin, pp. 43–63.

Brauner, C.J., Rummer, J.L., 2011. Gas transport and exchange: interaction between O<sub>2</sub> and CO<sub>2</sub> exchange. In: Farrell, A.P. (Ed.), *Encyclopedia of Fish Physiology: From Genome to Environment*. vol. 2. Academic Press, San Diego, pp. 916–920.

Brauner, C.J., Val, A.L., 1996. The interaction between O<sub>2</sub> and CO<sub>2</sub> exchange in the obligate air breather, *Arapaima gigas*, and the facultative air breather, *Lipossarcus pardalis*. In: Val, A.L., Almeida-Val, V.M.F., Randall, D.J. (Eds.), *Physiology and Biochemistry of the Fishes of the Amazon*. INPA, pp. 1029–1136.

Brauner, C.J., Wang, T., Wang, Y., Richards, J.G., Gonzalez, R.J., Bernier, N.J., Xi, W., Patrick, M., Val, A.L., 2004. Limited extracellular but complete intracellular acid-base regulation during short-term environmental hypercapnia in the armoured catfish, *Liposarcus pardalis*. *J. Exp. Biol.* 207, 3381–3390. <https://doi.org/10.1242/jeb.01144>.

Brauner, C.J., Shartau, R.B., Damsgaard, C., Esbbaugh, A.J., Wilson, R.W., Grosell, M., 2019. Acid[HYPHEN]base physiology and CO<sub>2</sub> homeostasis: regulation and compensation in response to elevated environmental CO<sub>2</sub>. In: Grosell, M., Munday, P.L., Farrell, A.P., Brauner, C.J. (Eds.), *Carbon Dioxide. Fish Physiology*. vol. 37. Academic Press, San Diego.

Bromhead, D., Scholey, V., Nicol, S., Margulies, D., Wexler, J., Stein, M., Hoyle, S., Lennert-Cody, C., Williamson, J., Havenhand, J., Ilyina, T., Lehodey, P., 2015. The potential impact of ocean acidification upon eggs and larvae of yellowfin tuna (*Thunnus albacares*). Deep Sea Res. II Top. Stud. Oceanogr. 113, 268–279. <https://doi.org/10.1016/j.dsro.2014.03.019>.

Caldeira, K., Wickett, M.E., 2003. Anthropogenic carbon and ocean pH. Nature 425, 365. <https://doi.org/10.1038/425365a>.

Calosi, P., Melatunian, S., Turner, L.M., Artioli, Y., Davidson, R.L., Byrne, J.J., Viant, M.R., Widdicombe, S., Rundle, S.D., 2017. Regional adaptation defines sensitivity to future ocean acidification. Nat. Commun. 8, 13994. <https://doi.org/10.1038/ncomms13994>

Cattano, C., Claudet, J., Domenici, P., Milazzo, M., 2018. Living in a high CO<sub>2</sub> world: a global meta-analysis shows multiple trait-mediated responses of fish to ocean acidification. Ecol. Monogr. 88, 320–335. <https://doi.org/10.1002/ecm.1297>.

Chambers, R.C., Candelmo, A.C., Habeck, E.A., Poach, M.E., Wieczorek, D., Cooper, K.R., Greenfield, C.E., Phelan, B.A., 2014. Effects of elevated CO<sub>2</sub> in the early life stages of summer flounder, *Paralichthys dentatus*, and potential consequences of ocean acidification. Biogeosciences 11, 1613–1626. <https://doi.org/10.5194/bg-11-1613-2014>.

Charnov, E.L., Berrigan, D., 1991. Dimensionless numbers and the assembly rules for life histories. Philos. Trans. R. Soc. Lond. B Biol. Sci. 332, 41–48. <https://doi.org/10.1098/rstb.1991.0031>.

Chevin, L.-M., Lande, R., Mace, G.M., 2010. Adaptation, plasticity, and extinction in a changing environment: towards a predictive theory. PLoS Biol. 8, e1000357. <https://doi.org/10.1371/journal.pbio.1000357>

Clack, J.A., 2007. Devonian climate change, breathing, and the origin of the tetrapod stem group. Integr. Comp. Biol. 47, 510–523. <https://doi.org/10.1093/icb/icm055>.

Collins, M., Knutti, R., Arblaster, J., Dufresne, J.-L., Fichefet, T., Friedlingstein, P., et al., 2013. Long-term climate change: projections, commitments and irreversibility. In: Stocker, T.F., Qin, D., Plattner, G.-K., Tignor, M., Allen, S.K., Boschung, J. et al., (Eds.), Climate Change 2013: The Physical Science Basis. Contribution of Working Group I to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change. Cambridge University Press, Cambridge, UK/New York, NY, USA, pp. 1029–1136. <https://doi.org/10.1017/CBO9781107415324.024>.

Crozier, L.G., Hutchings, J.A., 2014. Plastic and evolutionary responses to climate change in fish. Evol. Appl. 7, 68–87. <https://doi.org/10.1111/eva.12135>.

Davis, B.E., Komoroske, L.M., Hansen, M.J., Poletto, J.B., Perry, E.N., 2018. Juvenile rock fish show resilience to CO<sub>2</sub>-acidification and hypoxia across multiple biological scales. Conserv. Physiol. 6, coy038. <https://doi.org/10.1093/conphys/coy038>

Deigweiler, K., Koschnick, N., Pörtner, H.-O., Lucassen, M., 2008. Acclimation of ion regulatory capacities in gills of marine fish under environmental hypercapnia. Am. J. Physiol. Regul. Integr. Comp. Physiol. 295, R1660–R1670. <https://doi.org/10.1152/ajpregu.90403.2008>.

Diamond, S.E., Martin, R.A., 2016. The interplay between plasticity and evolution in response to human-induced environmental change. F1000Res. 5, 2835. <https://doi.org/10.12688/f1000research.9731.1>. F1000 Faculty Rev.

Donelson, J.M., Munday, P.L., McCormick, M.I., Pitcher, C.R., 2012. Rapid transgenerational acclimation of a tropical reef fish to climate change. Nat. Clim. Chang. 2, 30–32. <https://doi.org/10.1038/nclimate1323>.

Donelson, J.M., Salinas, S., Munday, P.L., Shama, L.N.S., 2018. Transgenerational plasticity and climate change experiments: where do we go from here? Glob. Chang. Biol. 24, 13–34. <https://doi.org/10.1111/gcb.13903>.

Doney, S.C., Fabry, V.J., Feely, R.A., Kleypas, J.A., 2009. Ocean acidification: the other CO<sub>2</sub> problem. *Ann. Rev. Mar. Sci.* 1, 169–192. <https://doi.org/10.1146/annurev.marine.010908.163834>.

Ellis, R.P., Urbina, M.A., Wilson, R.W., 2017. Lessons from two high CO<sub>2</sub> worlds—future oceans and intensive aquaculture. *Glob. Chang. Biol.* 23, 2141–2148. <https://doi.org/10.1111/gcb.13515>.

Enochs, I.C., Manzello, D.P., Donham, E.M., Kolodziej, G., Okano, R., Johnston, L., Young, C., Iguel, J., Edwards, C.B., Fox, M.D., Valentino, L., Johnson, S., Benavente, D., Clark, S.J., Carlton, R., Burton, T., Eynaud, Y., Price, N.N., 2015. Shift from coral to macroalgae dominance on a volcanically acidified reef. *Nat. Clim. Chang.* 5, 1083–1088. <https://doi.org/10.1038/nclimate2758>.

Erwin, D.H., Laflamme, M., Tweedt, S.M., Sperling, E.A., Pisani, D., Peterson, K.J., 2011. The cambrian conundrum: early divergence and later ecological success in the early history of animals. *Science* 334, 1091–1097. <https://doi.org/10.1126/science.1206375>.

Fabricius, K.E., Langdon, C., Uthicke, S., Humphrey, C., Noonan, S., De'ath, G., Okazaki, R., Muehllehner, N., Glas, M.S., Lough, J.M., 2011. Losers and winners in coral reefs acclimatized to elevated carbon dioxide concentrations. *Nat. Clim. Chang.* 1, 165–169.

Falconer, D.S., Mackay, T.F.C., 1996. *Introduction to Quantitative Genetics*, fourth ed. Longman, Essex.

Feely, R.A., Sabine, C.L., Hernandez-Ayon, J.M., Ianson, D., Hales, B., 2008. Evidence for upwelling of corrosive "acidified" water onto the continental shelf. *Science* 320, 1490–1492.

Hamilton, T.J., Logan, C.A., Fennie, H.W., Sogard, S.M., Barry, J.P., Makukhov, A.D., Tobosa, L.R., Boyer, K., Lovera, C.F., Bernardi, G., 2017. Species-specific responses of juvenile rockfish to elevated pCO<sub>2</sub>: from behavior to genomics. *PLoS One* 12, e0169670. <https://doi.org/10.1371/journal.pone.0169670>.

Hannan, K.D., Rummel, J.L., 2018. Aquatic acidification: a mechanism underpinning maintained oxygen transport and performance in fish experiencing elevated carbon dioxide conditions. *J. Exp. Biol.* 221, jeb154559. <https://doi.org/10.1242/jeb.154559>.

Heisler, N., 1982. Intracellular and extracellular acid-base regulation in the tropical fresh-water teleost fish *Synbranchus marmoratus* in response to the transition from water breathing to air breathing. *J. Exp. Biol.* 99, 9–28.

Heisler, N., 1986. Acid-base regulation in fishes. In: Heisler, N. (Ed.), *Acid-Base Regulation in Animals*. Elsevier, Amsterdam, pp. 309–356.

Hendry, A.P., Kinnison, M.T., Heino, M., Day, T., Smith, T.B., Fitt, G., Bergstrom, C.T., Oakeshott, J., Jørgensen, P.S., Zalucki, M.P., Gilchrist, G., Southerton, S., Sih, A., Strauss, S., Denison, R.F., Carroll, S.P., 2011. Evolutionary principles and their practical application. *Evol. Appl.* 4, 159–183. <https://doi.org/10.1111/j.1752-4571.2010.00165.x>.

Heuer, R.M., Grosell, M., 2014. Physiological impacts of elevated carbon dioxide and ocean acidification on fish. *Am. J. Physiol. Regul. Integr. Comp. Physiol.* 307, R1061–R1084. <https://doi.org/10.1152/ajpregu.00064.2014>.

Hoffmann, A.A., Sgro, C.M., 2011. Climate change and evolutionary adaptation. *Nature* 470, 479–485. <https://doi.org/10.1038/nature09670>.

Ishimatsu, A., Hayashi, M., Kikkawa, T., 2008. Fishes in high-CO<sub>2</sub>, acidified oceans. *Mar. Ecol. Prog. Ser.* 373, 295–302. <https://doi.org/10.3354/meps07823>.

Jarrold, M.D., Munday, P.L., 2019. Diel CO<sub>2</sub> cycles and parental effects have similar benefits to growth of a coral reef fish under ocean acidification. *Biol. Lett.* 15, 20180724. <https://doi.org/10.1098/rsbl.2018.0724>

Jensen, F., Nikinmaa, M., Weber, R., 1993. Environmental perturbations of oxygen transport in teleost fishes: causes, consequences and compensations. In: Rankin, J.C., Jensen, F. (Eds.), *Fish Ecophysiology*. vol. 9. Springer, Netherlands, pp. 161–179.

Johnson, D.W., Christie, M.R., Moye, J., 2010. Quantifying evolutionary potential of marine fish larvae: heritability, selection, and evolutionary constraints. *Evolution* 64, 2614–2628. <https://doi.org/10.1111/j.1558-5646.2010.01027.x>.

Jones, G.P., McCormick, M.I., Srinivasan, M., Eagle, J.V., 2004. Coral decline threatens fish biodiversity in marine reserves. *Proc. Natl. Acad. Sci. U. S. A.* 101, 8251–8253.

Junk, W.J., Soares, M.G.M., Bayley, P.B., 2007. Freshwater fishes of the Amazon River basin: their biodiversity, fisheries, and habitats. *Aquat. Ecosyst. Health Manage.* 10, 153–173. <https://doi.org/10.1080/14634980701351023>.

Kardong, K.V., 2012. *Vertebrates: Comparative Anatomy, Function, Evolution*, sixth ed. McGraw-Hill, New York.

Kelly, M.W., Padilla-Gamiño, J.L., Hofmann, G.E., 2013. Natural variation and the capacity to adapt to ocean acidification in the keystone sea urchin *Strongylocentrotus purpuratus*. *Glob. Chang. Biol.* 19, 2536–2546. <https://doi.org/10.1111/gcb.12251>.

Klironomos, F.D., Berg, J., Collins, S., 2013. How epigenetic mutations can affect genetic evolution: model and mechanism. *Bioessays* 35, 571–578. <https://doi.org/10.1002/bies.201200169>.

Kruuk, L.E.B., 2004. Estimating genetic parameters in natural populations using the “animal model” *Philos. Trans. R. Soc. Lond. B Biol. Sci.* 359, 873–890. <https://doi.org/10.1098/rstb.2003.1437>.

Laubenstein, T.D., Rummer, J.L., McCormick, M.I., Munday, P.L., 2019. A negative correlation between behavioural and physiological performance under ocean acidification and warming. *Sci. Rep.* 9, 4265. <https://doi.org/10.1038/s41598-018-36747-9>.

Lefevre, S., 2016. Are global warming and ocean acidification conspiring against marine ectotherms? A meta-analysis of the respiratory effects of elevated temperature, high CO<sub>2</sub> and their interaction. *Conserv. Physiol.* 4, cow009. <https://doi.org/10.1093/conphys/cow009>.

Lefevre, S., 2019. Effects of high CO<sub>2</sub> on oxygen consumption rates, aerobic scope and swimming performance. In: Grosell, M., Munday, P.L., Farrell, A.P., Brauner, C.J. (Eds.), *Carbon Dioxide. Fish Physiology*. vol. 37. Academic Press, San Diego.

Lohbeck, K.T., Riebesell, U., Reusch, T.B.H., 2012. Adaptive evolution of a key phytoplankton species to ocean acidification. *Nat. Geosci.* 5, 346–351. <https://doi.org/10.1038/ngeo1441>.

Lopes, A.F., Morais, P., Pimentel, M., Rosa, R., Munday, P.L., Goncalves, E.J., Faria, A.M., 2016. Behavioural lateralization and shoaling cohesion of fish larvae altered under ocean acidification. *Mar. Biol.* 163, 243. <https://doi.org/10.1007/s00227-016-3026-4>.

Lynch, M., Walsh, B., 1998. *Genetics and Analysis of Quantitative Traits*. Sinauer Associates, Inc., MA, USA.

Malvezzi, A.J., Murray, C.S., Feldheim, K.A., Dibattista, J.D., Garant, D., Gobler, C.J., Chapman, D.D., Baumann, H., 2015. A quantitative genetic approach to assess the evolutionary potential of a coastal marine fish to ocean acidification. *Evol. Appl.* 8, 352–362. <https://doi.org/10.1111/eva.12248>.

Marshall, D., Allen, R., Crean, A., 2008. The ecological and evolutionary importance of maternal effects in the sea. *Oceanogr. Mar. Biol. Annu. Rev.* 46, 203–250. <https://doi.org/10.1201/9781420065756.ch5>.

Märss, T., Gagnier, P.-Y., 2006. A new chondrichthyan from the Wenlock, Lower Silurian, of Bailleul-Hamilton Island, the Canadian Arctic. *J. Vertebr. Paleontol.* 21, 693–701. [https://doi.org/10.1671/0272-4634\(2001\)021\[0693:ancftw\]2.0.co;2](https://doi.org/10.1671/0272-4634(2001)021[0693:ancftw]2.0.co;2).

Masson-Delmotte, V., Schulz, M., Abe-Ouchi, A., et al., 2013. Information from paleoclimate archives. In: Stocker, T.F., Qin, D., Plattner, G.-K. et al., (Eds.), *Climate Change 2013: The Physical Science Basis. Contribution of Working Group I to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change*. Cambridge University Press, Cambridge, pp. 383–464.

McMahon, S.J., Donelson, J.M., Munday, P.L., 2018. Food ration does not influence the effect of elevated CO<sub>2</sub> on antipredator behaviour of a reef fish. *Mar. Ecol. Prog. Ser.* 586, 155–165. <https://doi.org/10.3354/meps12397>.

McNeil, B., Matsumoto, K., 2019. The changing ocean and freshwater CO<sub>2</sub> system. In: Grosell, M., Munday, P.L., Farrell, A.P., Brauner, C.J. (Eds.), *Carbon Dioxide. Fish Physiology*. vol. 37. Academic Press, San Diego.

Miller, G.M., Watson, S.-A., Donelson, J.M., McCormick, M.I., Munday, P.L., 2012. Parental environment mediates impacts of increased carbon dioxide on a coral reef fish. *Nat. Clim. Chang.* 2, 858–861. <https://doi.org/10.1038/nclimate1599>.

Mousseau, T.A., Roff, D.A., 1987. Natural selection and the heritability of fitness components. *Heredity* 59, 181–197.

Munday, P.L., 2014. Transgenerational acclimation of fishes to climate change and ocean acidification. *F1000 Prime Rep.* 6, 99. <https://doi.org/10.12703/P6-99>.

Munday, P.L., McCormick, M.I., Meekan, M., Dixson, D.L., Watson, S.-A., Chivers, D.P., Ferrari, M.C.O., 2012. Selective mortality associated with variation in CO<sub>2</sub> tolerance in a marine fish. *Ocean Acidif.* 1, 1–5. <https://doi.org/10.2478/oac-2012-0001>.

Munday, P.L., Pratchett, M.S., Dixson, D.L., Donelson, J.M., Endo, G.G.K., Reynolds, A.D., Knuckey, R., 2013a. Elevated CO<sub>2</sub> affects the behavior of an ecologically and economically important coral reef fish. *Mar. Biol.* 160, 2137–2144. <https://doi.org/10.1007/s00227-012-2111-6>.

Munday, P.L., Warner, R.R., Monro, K., Pandolfi, J.M., Marshall, D.J., 2013b. Predicting evolutionary responses to climate change in the sea. *Ecol. Lett.* 16, 1488–1500. <https://doi.org/10.1111/ele.12185>.

Munday, P.L., Cheal, A.J., Dixson, D.L., Rummer, J.L., Fabricius, K.E., 2014. Behavioural impairment in reef fishes caused by ocean acidification at CO<sub>2</sub> seeps. *Nat. Clim. Chang.* 4, 487–492. <https://doi.org/10.1038/nclimate2195>.

Munday, P.L., Donelson, J.M., Domingos, J.A., 2017. Potential for adaptation to climate change in a coral reef fish. *Glob. Chang. Biol.* 23, 307–317. <https://doi.org/10.1111/gcb.13419>.

Munday, P.L., Jarrold, M.D., Nagelkerken, I., 2019a. Ecological effects of elevated CO<sub>2</sub> on marine and freshwater fishes: from individual to community effects. In: Grosell, M., Munday, P.L., Farrell, A.P., Brauner, C.J. (Eds.), *Carbon Dioxide. Fish Physiology*. vol. 37. Academic Press, San Diego.

Munday, P.L., Schunter, C., Allan, B.J.M., Nicol, S., Parsons, D.M., Pether, S.M.J., Pope, S., Ravasi, T., Setiawan, A.N., Smith, N., Domingos, J.A., 2019b. Testing the adaptive potential of yellowtail kingfish to ocean warming and acidification. *Front. Ecol. Evol.* 7, 253. <https://doi.org/10.3389/fevo.2019.00253>.

Murray, C., Malvezzi, A., Gobler, C., Baumann, H., 2014. Offspring sensitivity to ocean acidification changes seasonally in a coastal marine fish. *Mar. Ecol. Prog. Ser.* 504, 1–11. <https://doi.org/10.3354/meps10791>.

Nagelkerken, I., Munday, P.L., 2016. Animal behaviour shapes the ecological effects of ocean acidification and warming: moving from individual to community-level responses. *Glob. Chang. Biol.* 22, 974–989. <https://doi.org/10.1111/gcb.13167>.

Near, T.J., Eytan, R.I., Dornburg, A., Kuhn, K.L., Moore, J.A., Davis, M.P., Wainwright, P.C., Friedman, M., Smith, W.L., 2012. Resolution of ray-finned fish phylogeny and timing of diversification. *Proc. Natl. Acad. Sci. U. S. A.* 109, 13698–13703. <https://doi.org/10.1073/pnas.1206625109>.

Parker, L.M., Ross, P.M., O'Connor, W.A., 2011. Populations of the Sydney rock oyster, *Saccostrea glomerata*, vary in response to ocean acidification. *Mar. Biol.* 158, 689–697. <https://doi.org/10.1007/s00227-010-1592-4>.

Patterson, C., 1993. Osteichthyes: Teleostei. In: Benton, M.J. (Ed.), *The Fossil Record 2*. Springer, pp. 621–656.

Pratchett, M.S., Munday, P.L., Wilson, S.K., Graham, N.A.J., Cinner, J.E., Bellwood, D.R., Jones, G.P., Polunin, N.V.C., McClanahan, T.R., 2008. Effects of climate-induced coral bleaching on coral-reef fishes ecological and economic consequences. *Oceanogr. Mar. Biol. Annu. Rev.* 46, 251–296. <https://doi.org/10.1201/9781420065756>.

Randall, D.J., Rummer, J.L., Wilson, J.M., Wang, S., Brauner, C.J., 2014. A unique mode of tissue oxygenation and the adaptive radiation of teleost fishes. *J. Exp. Biol.* 217, 1205–1214. <https://doi.org/10.1242/jeb.093526>.

Reusch, T.B.H., 2014. Climate change in the oceans: evolutionary versus phenotypically plastic responses of marine animals and plants. *Evol. Appl.* 7, 104–122. <https://doi.org/10.1111/eva.12109>.

Royer, D.L., Berner, R.A., Montañez, I.P., Tabor, N.J., Beerling, D.J., 2004. CO<sub>2</sub> as a primary driver of Phanerozoic climate. *GSAToday* 14, 4–10. [https://doi.org/10.1130/1052-5173\(2004\)014<4](https://doi.org/10.1130/1052-5173(2004)014<4).

Rummer, J.L., Brauner, C.J., 2011. Plasma-accessible carbonic anhydrase at the tissue of a teleost fish may greatly enhance oxygen delivery: *in vitro* evidence in rainbow trout, *Oncorhynchus mykiss*. *J. Exp. Biol.* 214, 2319–2328.

Rummer, J.L., Munday, P.L., 2017. Climate change and the evolution of reef fishes: past and future. *Fish Fish.* 18, 22–39. <https://doi.org/10.1111/faf.12164>.

Rummer, J.L., Mckenzie, D.J., Innocenti, A., Supuran, C.T., Brauner, C.J., 2013. Root effect hemoglobin may have evolved to enhance general tissue oxygen delivery. *Science* 340, 1327–1329. <https://doi.org/10.1126/science.1233692>.

Ryu, T., Veilleux, H.D., Donelson, J.M., Munday, P.L., Ravasi, T., 2018. The epigenetic landscape of transgenerational acclimation to ocean warming. *Nat. Clim. Chang.* 8, 504–509. <https://doi.org/10.1038/s41558-018-0159-0>.

Schade, F.M., Clemmesen, C., Mathias Wegner, K., 2014. Within- and transgenerational effects of ocean acidification on life history of marine three-spined stickleback (*Gasterosteus aculeatus*). *Mar. Biol.* 161, 1667–1676. <https://doi.org/10.1007/s00227-014-2450-6>.

Schmidt, M., Gerlach, G., Leo, E., Kunz, K.L., Swoboda, S., Pörtner, H.O., Bock, C., Storch, D., 2017. Impact of ocean warming and acidification on the behaviour of two co-occurring gadid species, *Boreogadus saida* and *Gadus morhua*, from Svalbard. *Mar. Ecol. Prog. Ser.* 571, 183–191. <https://doi.org/10.3354/meps12130>.

Schunter, C., Welch, M.J., Nilsson, G.E., Rummer, J.L., Munday, P.L., Ravasi, T., 2018. An interplay between plasticity and parental phenotype determines impacts of ocean acidification on a reef fish. *Nat. Ecol. Evol.* 2, 334–342. <https://doi.org/10.1038/s41559-017-0428-8>.

Skov, P.V., 2019. CO<sub>2</sub> in aquaculture. In: Grosell, M., Munday, P.L., Farrell, A.P., Brauner, C.J. (Eds.), *Carbon Dioxide. Fish Physiology*. vol. 37. Academic Press, San Diego.

Snyder, J.T., Murray, C.S., Baumann, H., 2018. Potential for maternal effects on offspring CO<sub>2</sub> sensitivities in the Atlantic silverside (*Menidia menidia*). *J. Exp. Mar. Biol. Ecol.* 499, 1–8. <https://doi.org/10.1016/j.jembe.2017.11.002>.

Stiasny, M.H., Mittermayer, F.H., Göttler, G., Bridges, C.R., Falk-Petersen, I.B., Puvanendran, V., Mortensen, A., Reusch, T.B.H., Clemmesen, C., 2018. Effects of parental acclimation and energy limitation in response to high CO<sub>2</sub> exposure in Atlantic cod. *Sci. Rep.* 8, 8348. <https://doi.org/10.1038/s41598-018-26711-y>.

Sunday, J.M., Crim, R.N., Harley, C.D.G., Hart, M.W., 2011. Quantifying rates of evolutionary adaptation in response to ocean acidification. *PLoS One* 6, e22881. <https://doi.org/10.1371/journal.pone.0022881>.

Sunday, J.M., Calosi, P., Dupont, S., Munday, P.L., Stillman, J.H., Reusch, T.B.H., 2014. Evolution in an acidifying ocean. *Trends Ecol. Evol.* 29, 117–125. <https://doi.org/10.1016/j.tree.2013.11.001>.

Sundin, J., Amcoff, M., Mateos-González, F., Raby, G.D., Jutfelt, F., Clark, T.D., 2017. Long-term exposure to elevated carbon dioxide does not alter activity levels of a coral reef fish in response to predator chemical cues. *Behav. Ecol. Sociobiol.* 71, 108. <https://doi.org/10.1007/s00265-017-2337-x>.

Tasoff, A.J., Johnson, D.W., 2019. Can larvae of a marine fish adapt to ocean acidification? Evaluating the evolutionary potential of California Grunion (*Leuresthes tenuis*). *Evol. Appl.* 12, 560–571. <https://doi.org/10.1111/eva.12739>.

Tresguerres, M., Milsom, W.K., Perry, S.F., 2019. CO<sub>2</sub> and acid-base sensing. In: Grosell, M., Munday, P.L., Farrell, A.P., Brauner, C.J. (Eds.), *Carbon Dioxide. Fish Physiology*. vol. 37. Academic Press, San Diego.

Ultsch, G.R., 1996. Gas exchange, hypercarbia and acid-base balance, paleoecology, and the evolutionary transition from water-breathing to air-breathing among vertebrates. *Palaeogeogr. Palaeoclimatol. Palaeoecol.* 123, 1–27.

Vargas, C.A., Lagos, N.A., Lardies, M.A., Duarte, C., Manríquez, P.H., Aguilera, V.M., Broitman, B., Widdicombe, S., Dupont, S., 2017. Species-specific responses to ocean acidification should account for local adaptation and adaptive plasticity. *Nat. Ecol. Evol.* 1, 0084. <https://doi.org/10.1038/s41559-017-0084>.

Ward, P., 2006. *Out of Thin Air: Dinosaurs, Birds, and Earth's Ancient Atmosphere*. Joseph Henry Press, Washington D.C.

Welch, M.J., Munday, P.L., 2017. Heritability of behavioural tolerance to high CO<sub>2</sub> in a coral reef fish is masked by nonadaptive phenotypic plasticity. *Evol. Appl.* 10, 682–693. <https://doi.org/10.1111/eva.12483>.

Welch, M.J., Watson, S.-A., Welsh, J.Q., McCormick, M.I., Munday, P.L., 2014. Effects of elevated CO<sub>2</sub> on fish behaviour undiminished by transgenerational acclimation. *Nat. Clim. Chang.* 4, 1086–1089. <https://doi.org/10.1038/nclimate2400>.

Wells, R.M.G., Baldwin, J., Ryder, J.M., 1992. Respiratory function and nucleotide composition of erythrocytes from tropical elasmobranchs. *Comp. Biochem. Physiol. A Physiol.* 103, 157–162. [https://doi.org/10.1016/0300-9629\(92\)90256-P](https://doi.org/10.1016/0300-9629(92)90256-P).

Wilson, A.J., Pemberton, J.M., Pilkington, J.G., Clutton-Brock, T.H., Coltman, D.W., Kruuk, L.E.B., 2007. Quantitative genetics of growth and cryptic evolution of body size in an island population. *Evol. Ecol.* 21, 337–356. <https://doi.org/10.1007/s10682-006-9106-z>.

Zeebe, R.E., Ridgwell, A., Zachos, J.C., 2016. Anthropogenic carbon release rate unprecedented during the past 66 million years. *Nat. Geosci.* 9, 325–329.

Zinsmeister, W.J., 1998. Discovery of fish mortality horizon at the K-T boundary on Seymour Island: re-evaluation of events at the end of the Cretaceous. *J. Paleo.* 72, 556–571.